

Impact of particulate matter on the morbidity and mortality and its assessment of economic costs

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Abstract. Kazakhstan's cities experience high concentrations levels of atmospheric particulate matter (PM), which is well-known for its highly detrimental effect on the human health. A further increase in PM concentrations in the future could lead to a higher air pollution-caused morbidity and mortality, causing an increase in healthcare expenditures by the government. However, to prevent elevated PM concentrations in the future, more stringent standards could be implemented by lowering current maximum allowable PM concentration limit to Organization for Economic Co-operation and Development (OECD)'s limits. Therefore, this study aims to find out what impact this change in environmental policy towards PM has on state economy in the long run. Future PM₁₀ and PM_{2.5} concentrations were estimated using multiple linear regression based on gross regional product (GRP) and population growth parameters. Dose-response model was based on World Health Organization's approach for the identification of mortality, morbidity and healthcare costs due to air pollution. Analysis of concentrations revealed that only 6 out of 21 cities of Kazakhstan did not exceed the EU limit on PM₁₀ concentration. Changing environmental standards resulted in the 71.7% decrease in mortality and 77% decrease in morbidity cases in all cities compared to the case without changes in environmental policy. Moreover, the cost of morbidity and mortality associated with air pollution decreased by \$669 million in 2030 and \$2183 million in 2050 in case of implementation of OECD standards. Thus, changing environmental regulations will be beneficial in terms of both of mortality reduction and state budget saving.

Keywords: air pollution; Kazakhstan; mortality and morbidity; healthcare cost; environmental policy

1. Introduction

Airborne particulate matter (PM), one of the main indicators of air pollution, consists of various solid and liquid particles such as elemental and organic carbon, nitrates, sulfates, organic compounds, biological compounds (e.g., endotoxin, cell fragments) and various metals (copper, zinc, iron, nickel and vanadium), which are suspended in the air. Particles can combine and form heterogeneous mixtures of various sizes (PM_{2.5} – with a diameter smaller than 2.5 μm; PM₁₀ – with

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diameter of 2.5 – 10 μm) and chemical composition (Kim *et al.* 2015). PM draws the attention of the experts worldwide because of its adverse effect on human health (Chiang *et al.* 2016, Li and Lin 2014). Potential health risks due to exposure to PM is directly related to its particle size. Recent research conducted by Jiang *et al.* (2018) evaluated health risk from collected PM in Zhengzhou, China. The study elucidated that crustal elements are mainly found in PM_{2.5-10}, while elements generated by anthropogenic sources (e.g., Pb, Zn, As, Cu, Cd, etc.) mainly exist in fine particles rather than in coarse particles (Jiang *et al.* 2018). The study conducted by Dappe *et al.* (2018) also evidences that finest fraction of PMs are the most Pd enriched particles collected from lead recycling plant, suggesting that smaller sized particles are more enriched with elements from anthropogenic sources. PM in the atmosphere could lead to a more dangerous human health consequence compared to other common air pollutants such as carbon monoxide or ground-level ozone (Kim *et al.* 2015). PM₁₀ and PM_{2.5} coming from oil refineries results in a health effects such as impaired lungs function, aggravates asthma, respiratory diseases, heart attacks. Also, long-term and short-term exposure causes mortalities associated with chronic obstructive pulmonary disease (COPD) and ischemic heart disease (IHD) in population older than 30 years old (Amoatey *et al.* 2019). A systematic review by Yang *et al.* (2019) concludes that black carbon and organic carbon along with nitrate, sulfate, ammonium, Fe, Si, V and Zn on a surface of PM_{2.5} cause adverse health effects. Epidemiologic studies have revealed a strong positive correlation between the level of PM₁₀ and numbers of hospital admissions due to respiratory and cardiovascular disorders all over the world (Anderson *et al.* 1997, Burnett *et al.* 1999, Pope *et al.* 1995, Zanobetti *et al.* 2000). Major negative effects of exposure to PM_{2.5} include irritation of the lung airways, difficult breath, or coughing and premature death due to heart or lung disease, irregular heartbeat, decreased lung function and nonfatal heart attacks (Correia *et al.* 2015, Atkinson *et al.* 2010, Cadelis *et al.* 2014, Fang *et al.* 2013, Meister *et al.* 2012). Air pollution directly increases in the number of deaths by more than two million each year globally due to the negative effect on the lungs and the respiratory system (Shah *et al.* 2013). Among those deaths, cases related to death caused by PM_{2.5} are around 2.1 million (Chuang *et al.* 2011, Shah *et al.* 2013).

The exposure to PM of various particle sizes can lead to a wide range of human health disorders, which in turn cause economical losses. Adverse health effects are associated with additional health expenditure, labor productivity loss, and work time loss and consequently could lead to a significant impact on the economy (Wu *et al.* 2017). Previous scientific works successfully quantified the economic impact due to exposure to outdoor air pollution (Hunt *et al.* 2016). For example, according to the study conducted in the United States, the productivity of agricultural workers was significantly influenced by the ozone levels which were not in the range of federal air-quality standards (Graff Zivin and Neidell 2011). Also, a study conducted in Mexico City states that work hours per week were increased by 1.3 h (3.5%) when SO₂ levels were decreased by 19.7% (Hanna and Oliva 2015). In China, air pollution leads to economic losses equal to 0.72-6.94% of the regional GDP of the cities and provinces (Huang and Zhang 2013, Huang *et al.* 2012, Kan and Chen 2004, Wenbo and Shiqiu 2010, Zhang *et al.* 2010). In 2005, global welfare loss due to outdoor air pollution was more than \$5 trillion, including \$1.7 trillion in OECD countries, \$0.5 trillion in India, and \$1.4 trillion in China (OECD 2014). According to OECD's computable general equilibrium model ENV-Linkages, the estimated effect of outdoor air pollution will lead to economic losses equal to 1% of global GDP by 2060 (OECD 2016). In this regard, it is necessary to thoroughly elucidate the effect of outdoor air pollution on health and its economic impact.

Kazakhstan is ninth largest country by area in the world and the second largest residential coal

Table 1 Calculated average annual concentration of PM₁₀ for all cities of Kazakhstan in 2011-2017, µg/m³ (Kazhydromet 2018)

Cities/year	2011	2012	2013	2014	2015	2016	2017
Nur-Sultan	81	80	82	71	99	180	135
Zhezkazgan	131	129	137	148	127	135	144
Temirtau	129	107	95	135	136	135	122
Shymkent	112	94	103	108	105	90	115
Balkhash	68	78	79	91	91	90	79
Almaty	60	73	52	63	80	90	77
Aktau	106	119	126	105	83	45	95
Atyrau	83	296	197	110	85	45	62
Taraz	47	141	159	47	45	100	40
Pavlodar	92	91	82	80	99	45	41
Ekibastuz	N/A	29	28	43	74	45	58
Karaganda	26	29	45	48	53	45	63
Semey	59	35	27	46	56	45	54
Ust-Kamenogorsk	68	78	87	64	45	45	45
Taldykorgan	N/A	53	45	54	52	45	29
Kokshetau	N/A	0	0	8	68	18	135
Petropavlovsk	N/A	37	35	36	38	45	0
Uralsk	32	N/A	N/A	27	63	40	20
Kyzylorda	N/A	8	40	9	18	0	27
Aktobe	N/A	N/A	25	14	14	9	14
Kostanay*	N/A	9	0	49	85	50	10

* For Kostanay, measured PM concentration was used instead of calculated

consumer per capita in the world in 2014 (Torkmahalleh *et al.* 2020). The climate of Kazakhstan is dry and major lands of the country are steppes and deserts. The average annual precipitation is higher than 300 mm in northern areas, while central and southern parts is estimated to have only 150-200 mm of average annual precipitation. The temperature in winter can decrease up to -45°C (with average of -4 to -9°C), while in summer it can go up to 45°C (with average of +19 to +26°C) (Akhanova *et al.* 2020). Also, research study in Nur-Sultan, the capital of Kazakhstan, states that PM₁ is found to be the dominant PM fraction (77-94%) (Torkmahalleh *et al.* 2020). The major source of PM in Kazakhstan is not identified yet, and source apportionment with chemical analysis of PM is required (Kerimray *et al.* 2020). However, cheap fossil fuel, especially coal, consumption for heating purposes might explain high PM concentrations during winter seasons (Vinnikov *et al.* 2020). Serious air pollution cases and its impact on people in Kazakhstan were reported by several studies (Kenessary *et al.* 2019, Kerimray *et al.* 2020). Kenessariyev *et al.* (2013) estimated the costs associated with air pollution in Kazakhstan, which was based on the methodology of the World Health Organization, which suggested using the log-linear approximation of the health risk function (Ostro 2004). It was reported that mortality rates due to air pollution in Kazakhstan are very high, several times higher than the number that was estimated

Table 2 The population of cities in Kazakhstan (Committee on Statistics 2018)

City	Population	Population aged 0-15	Population aged 16-61	Population aged 63+	Total mortality rate per 100,000 people
Nur-Sultan	1030577	302021	652339	76217	626.3
Zhezkazgan	91045	25702	53510	11833	861.8
Temirtau	186003	41048	113829	31126	861.8
Shymkent	952170	330914	547416	73840	811.5
Balkhash	78722	20047	47912	10763	861.8
Almaty	1801993	413477	1170931	217585	575.3
Aktau	186238	59962	107589	18687	820.4
Atyrau	327852	113809	187545	26498	896.3
Taraz	355825	119277	197282	39266	850.9
Pavlodar	360048	80248	221302	58498	790.9
Ekibastuz	152853	35768	97824	19261	790.9
Karaganda	501419	118898	306266	76255	861.8
Semey	347284	81852	216866	48566	799.1
Ust-Kamenogorsk	341064	70238	210445	60381	799.1
Taldykorgan	171726	48603	102028	21095	887.5
Kokshetau	159807	41935	96744	21128	951.1
Petropavlovsk	218031	44454	133405	40172	813.7
Uralsk	303971	80041	183488	40442	834.9
Kyzylorda	294415	105732	164510	24173	913.3
Aktobe	477052	135429	293252	48371	753.4
Kostanay	239652	51086	151873	36693	782.7

by the WHO and reported in “Country profile for the ecological burden of disease” (WHO 2009). While WHO states in their report for Kazakhstan entitled “Country Profile for the Environmental Burden of Disease” that the mean urban PM₁₀ concentration in Kazakhstan is 25 µg/m³, Kenessariyev *et al.* (2013) found that the actual concentrations are much higher and are dependent on the pollution characteristics of the city. In 2014, Brody and Golub (2014) reported that the average concentrations of PM_{2.5} in the cities of Almaty and Nur-Sultan are 5 times higher than the standard set by WHO. They also recommended to implement risk assessment procedures, assess costs and benefits, and cost-effectiveness, and that these analytical procedures should be part of the decision-making process to achieve standards (Brody and Golub 2014). Moreover, a high carcinogenic and non-carcinogenic risks due to air pollution with suspended particles, oxides and dioxides of nitrogen and sulfur were observed in most of Kazakhstan’s cities (Kenessary *et al.* 2019). Stricter air guideline values could be implemented to reduce the adverse impact of air pollution on the population, which in turn is associated with economic loss. Implementing air quality standards of developed countries such as OECD air guidelines could help to mitigate the effect of air pollution on people and consequently reduce healthcare-related governmental expenses.

Quantitative assessment of the outcomes of different environmental policies and evaluation of

Table 3 Multiple linear regression coefficients

City	R-square	Y-intersection (PM _{2.5})	Variable X ₁ (GRP)	Variable X ₂ (Population)
Nur-Sultan	0.987	0.738	0.301	-0.029
Zhezkazgan	0.978	0.694	0.308	0.000
Temirtau	0.981	-0.450	0.454	0.998
Shymkent	0.989	0.979	0.019	0.002
Balkhash	0.978	0.129	0.843	0.034
Almaty	0.979	0.622	0.085	0.295
Aktau	0.986	2.522	0.089	-1.611
Atyrau	0.945	1.202	0.108	-0.316
Taraz	0.983	1.841	-0.376	-0.460
Pavlodar	0.994	3.933	-0.475	-2.459
Ekibastuz	0.756	-0.926	2.391	-0.470
Karaganda	0.944	-3.380	0.390	4.040
Semipalatinsk	0.987	-2.716	-0.014	3.727
Ust-Kamenogorsk	0.994	1.495	-0.168	-0.327
Taldykorgan	0.999	1.396	-0.020	-0.377
Kokshetau	0.959	-3.304	0.641	3.685
Petropavlovsk	0.872	-2.673	0.043	3.694
Uralsk	0.960	1.117	-0.113	-0.003
Kyzylorda	1.000	1.704	-0.059	-0.645
Aktobe	0.990	1.186	-0.029	-0.157
Kostanai	0.992	-1.907	0.109	2.816

costs and advantages should be implemented for the decision making regarding environmental policies in Kazakhstan. In this regard, two scenarios were assumed: “business as usual” (no change in standards) and “OECD standards” (implementing OECD standards). Thus, the study seeks to determine economic benefits from a change in the environmental policy of Kazakhstan in the long run. In particular, objectives are (1) to investigate expected values of the gross regional product (GRP) and PM concentrations in the atmosphere of Kazakhstan cities until 2050, and (2) to elucidate and compare the impact of emitted air pollutant on premature deaths, diseases, health care costs and work capacity in both scenarios.

2. Materials and methods

2.1 Study area and data source

The study area covers all major cities of Kazakhstan including Nur-Sultan, Zhezkazgan, Temirtau, Shymkent, Balkhash, Almaty, Aktau, Atyrau, Taraz, Pavlodar, Ekibastuz, Karaganda, Semipalatinsk, Ust-Kamenogorsk, Taldykorgan, Kokshetau, Petropavlovsk, Uralsk, Kyzylorda, Aktobe, Kostanay (Fig. S1). Data on concentrations of PM from 2011 to 2017 were taken from

information bulletins on the state of the environment of Regional State Enterprise (RSE) “Kazhydromet” (Table 1) (Kazhydromet 2018). RSE “Kazhydromet” continuously measured the concentration of total suspended particles (TSP) and PM_{10} at national air monitoring network’s observation posts comprising manually controllable and automatic stations. Measurement was conducted based on gravimetric and light scattering approaches. All locations of observation posts are listed in the Table S1. Due to the unavailability of measurements, PM_{10} concentrations were determined based on the ratio of $TSP/PM_{10} = 0.45$, following the work by the World Bank (World Bank 2012). Table 2 demonstrates demographic data (population and mortality rates) used for estimating future concentrations (Committee on Statistics 2018).

2.2 Statistical analysis

Future concentrations of air pollutants were estimated using multiple regression analysis in Excel 2016 software. Hence, correlation coefficients of the regression equation were obtained for the dependence of the growth index of PM_{10} concentration on growth indices of GRP and population. The growth index is defined as the values of some parameter (concentration, population or GRP) in the current year divided by the values of the same parameter in a reference year (2011). Data on growth indices of the GRP and population were obtained from the Committee on Statistics of the Ministry of National Economy. Since lack of data on GRP of cities separately, the GRP growth index of the regions, where the particular city is located, were assumed to be equal to the GRP growth index in the city (Committee on Statistics 2018) (Figs. S2 and S3). Calculation of the growth index of PM_{10} was reported by (Kerimray *et al.* 2018). General statistical equations by the method of least squares were used to calculate the coefficients based on these 2011-2017 data. The multiple linear regression coefficients for $PM_{2.5}$ (y), GRP (X_1) and population (X_2) are presented in Table 3.

2.3 Assessment of mortality and morbidity

Assessment of mortality was conducted using the estimated concentrations of PM_{10} using “concentration-response” functions developed by the World Health Organization were used in this study (Ostro 2004). The method used in this study estimates the impact of PM on population (long-term exposure of adults to $PM_{2.5}$ associated with cardiopulmonary mortality & lung cancer and short-term exposure of children to PM_{10} associated with the respiratory mortality) via linear “concentration-response” function. It is worth pointing out that the original methodology guideline was designed for estimation of mortality for adults > 30 years old and children < 5 years old. However, due to the availability of demographic data, the exposed population was assumed to include adults (> 18 years old). For children, only demographic data for the age group of 0-15 years old was available. Therefore, it was assumed that the population of children aged 0-5 years old is the third (33%) of the population of all children (0-15 years old), who are living in urban areas. The average percent of children aged 0-15 living in cities is 26.8% of the total urban population based on a comparison of all cities. Thus, according to our assumption, children living in urban areas and aged < 5 years old comprise about 8.8% of the urban population (0.33×0.268). This correlates well with existing studies because the percentage of children aged < 5 years old in urban areas was reported to be ~ 8.2% in developing countries (UNDESA 2017).

The linear relationship between the concentration of air pollutants and associated health effects are demonstrated in Eq. (1). RR is the relative risk. X_0 is a target concentration of pollutants and is

Table 4 Methodology for calculating the relative risk of mortality (Ostro 2004)

Parameter	β (lower-upper bound)	DALY/10000 cases
Mortality from cardiopulmonary diseases for adults (long-term)	0.00893 (0.00322-0.01464)	80000
Mortality from lung cancer for adults (long-term)	0.01267 (0.00432-0.02102)	80000
Mortality of acute respiratory diseases of the lower respiratory tract (ARI NDP) of children under 5 years (short-term)	0.00166 (0.00034-0.0030)	340000

assumed to be $15 \mu\text{g}/\text{m}^3$ for PM_{10} (the lowest PM_{10} level observed in rural areas of Kazakhstan) and $7.5 \mu\text{g}/\text{m}^3$ for $\text{PM}_{2.5}$ (World Bank 2012). The target concentration is used to estimate the overall damage of human health by atmospheric pollution and represents the lowest level that could be achieved by proper policy measures. X is the existing concentration of an air pollutant. β is a concentration-response coefficient, i.e., environmental burden of disease representing the increase in the relevant health effect due to the increase in air pollutant's concentration by $1 \mu\text{g}/\text{m}^3$. Table 4 demonstrates the details of the calculation of relative risks of mortality from air pollution-related diseases.

Due to the unavailability of statistical data on $\text{PM}_{2.5}$, the $\text{PM}_{2.5}/\text{PM}_{10}$ ratio was assumed to convert PM_{10} data to $\text{PM}_{2.5}$ data. Ostro (2004) used a ratio of 0.5 for developing countries (Ostro 2004). Different studies demonstrate a wide range of ratios from 0.31 in Bahrain to 0.7 in northern China (Coskuner *et al.* 2018, Duan *et al.* 2015, Zhao *et al.* 2019). The ratio was heavily affected by local meteorological conditions. Stable atmospheric conditions are associated with a higher ratio (Xu *et al.* 2017). Kenessariyev *et al.* (2013) assumed ratios of 0.2-0.3 for steppe areas, which are likely to have sandstorms, and 0.4 for areas with a low probability of sandstorms in Kazakhstan (Kenessariyev *et al.* 2013). In this study, a rough assumption of $\text{PM}_{2.5}/\text{PM}_{10}$ ratio = 0.45 was accepted because the objective of the study was not to estimate the mortality and morbidity but to estimate the difference in health-related costs between the 2 scenarios.

$$RR = \exp[\beta(X - X_0)] \quad (1)$$

The mortality was calculated using Eq. (2). E is the number of deaths, and AF is related to an attributable fraction and was calculated with Eq. (3). B is the overall mortality rate related to a relevant health effect and P is the exposed population (Ostro 2004). Data for mortality from a specific health effect (cardiopulmonary mortality, lung cancer and acute lower respiratory infection-related total mortality) was not available. Therefore, we used the method of World Bank study to estimate specific mortality. The percentage of cardiovascular and lung cancer mortality in adults was assumed to be 35.5% and 2% of total mortality, respectively. For children under 5 years old, mortality from acute lower respiratory infections was assumed to be 6.8% of total mortality.

$$E = AF \times B \times P \quad (2)$$

$$AF = (RR - 1) / RR \quad (3)$$

The morbidity was estimated by calculating the number of chronic bronchitis cases, the number of hospitalizations, appeals for urgent medical help, days of limited activity, diseases of the lower respiratory ways in children and symptoms of respiratory diseases. For the assessment of the relationship of morbidity incidence to PM_{10} , its response coefficients by the World Bank were used

Table 5 Coefficients of reaction to the impact of urban atmospheric pollution, to calculate the incidence rate (World Bank 2012)

Impact on health PM ₁₀	Unit	Impact on 1 µg/m ³
Chronic bronchitis	100000 adults	0.9
Number of hospitalizations	100000 population	1.2
Appeals for urgent medical help	100000 population	23.5
Days of limited activity	100000 adults	5750
Diseases of the lower respiratory ways in children	100000 children	169
Symptoms of respiratory diseases	100000 adults	18300

Table 6 The cost of a unit of medical care and temporary losses associated with disease (World Bank 2012)

Costs due to illness	Unit	Cost of one unit (USD)
Chronic bronchitis (PM ₁₀)	Day	14620
Hospitalization (PM ₁₀)	Visit	587
Requests for emergency medical care (PM ₁₀)	Visit	79
Days of limited activity (PM ₁₀)	Day	3.4
Diseases of the lower respiratory tract in children (PM ₁₀)	Day	63
Symptoms of respiratory diseases (PM ₁₀)	Day	0.8

based on the analysis of international studies (Table 5).

2.4 Calculations of health care costs

DALY is a disability-adjusted life year, i.e., the sum of years lost due to premature death or disability (World Bank 2012). Table 4 also represents the values of DALY per 10000 cases of health effects that were obtained from the World Bank study. The monetary effect of mortality can be estimated using Eq. (4).

$$\text{Monetary effect} = \text{Gross Domestic Product} \times \text{DALY} \quad (4)$$

Morbidity-related costs, i.e., health care costs were calculated using the World Bank approach which includes the adapted approach to human capital, the value of statistical life, and the cost of illness. Table 6 shows the costs associated with the treatment of diseases and the loss of time due to the disease, which is based on data on medical costs received from the Ministry of Health of the Republic of Kazakhstan and information on wages in Kazakhstan and calculated by the World Bank experts (World Bank 2012).

3. Results and discussion

3.1 Estimation of concentrations of PM_{2.5} and PM₁₀ until 2050

In the “Business as usual” scenario, the concentrations until 2050 followed the same trend as those from 2011 to 2017 (mostly growing), while the “OECD” scenario’s estimated for future

Table 7 Comparison of Kazakhstani standard values with the World Health Organization and the European Union standards (European Commission 2019, Kazhydromet 2018, WHO 2018)

Pollutant	Concentration, $\mu\text{g}/\text{m}^3$			Time interval
	WHO	EU	Kazakhstan	
PM ₁₀	20	40	300 (max. single concentration)	1 year
	50	50	60	Average daily
PM _{2.5}	10	25	160 (max. single concentration)	1 year
	25		35	Average daily

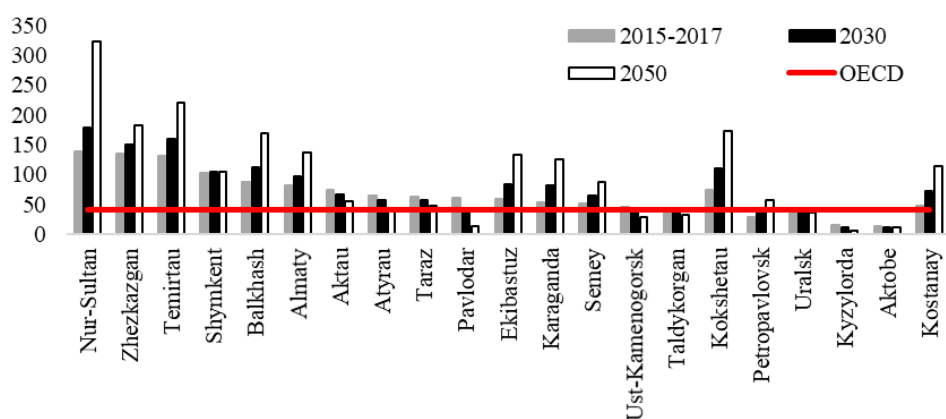


Fig. 1 Predicted concentrations of PM₁₀ in the air up to 2050 according to the “Business as usual” scenario, $\mu\text{g}/\text{m}^3$

PM₁₀ and PM_{2.5} concentrations were calculated based on the guideline values for the air pollutants of EU. “OECD” scenario assumed that, by the year 2030, the air pollutant level will reach the annual limit value set in the EU standard ($40 \mu\text{g}/\text{m}^3$, Table 7). Table S2 demonstrates the predicted concentrations for the “OECD” scenario. The projection was done for years 2017 (initial year), 2020, 2025, 2030, 2035, 2040 and 2050. Concentrations in the “OECD” scenario was calculated based on assumption that the initial concentration of PM decreases by the same amount each period until 2030 (there are 3 periods, i.e., 2017-2020, 2020-2025, 2025-2030). The average concentration in 2015-2017 in Nur-Sultan ($138 \mu\text{g}/\text{m}^3$) for an example was assumed to decrease by $32.1 \mu\text{g}/\text{m}^3$ each year until 2030 because $(138 - 40)/3 = 32.1 \mu\text{g}/\text{m}^3$. It is worth pointing out that if concentrations decreased in the “Business as usual” scenario (Pavlodar, Ust-Kamenogorsk, Taldykorgan, Uralsk, Kyzylorda and Aktope), the same trend was assumed in “OECD scenario”, i.e., predicted concentrations were left as they are in “OECD” scenario if they were $< 40 \mu\text{g}/\text{m}^3$ to prevent overestimation of results.

To calculate PM₁₀ concentrations until 2050 with no amendment in air guidelines, estimations of future populations and GRP values were needed. The estimated future population of each city and GRP until 2050 can be seen in Figs. S4 and S5, respectively. The future population and GRP were assumed to grow at an exponential rate. Population-projected growth rates were high in Shymkent, Nur-Sultan, Almaty, Atyrau and Uralsk because of the migration of citizens (Olzhaev 2014, Userbayeva 2019, Zhusupova and Kenesov 2012). Fig. S5 shows that Almaty and Nur-Sultan have the largest increase in GRP because the business activity and finance are centered in

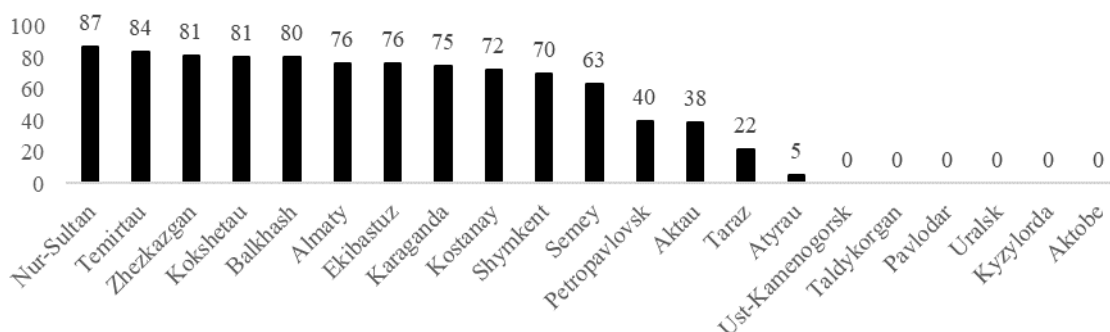


Fig. 2 Mortality reduction in 2050 in the scenario “OECD Standards” in comparison to the “Business as usual” scenario in %

Almaty, while Nur-Sultan is the capital of the country with a fast-growing population and migration rate into the city.

The estimated future concentrations of PM_{10} and $PM_{2.5}$ for the “Business as usual” scenario were demonstrated in Fig. 1. The estimated future $PM_{2.5}$ concentrations were demonstrated in Table S3. The air pollutants’ concentrations were exceeded OECD’s annual limit value for PM_{10} by 2050 in all cities, except Kyzylorda ($5.9 \mu\text{g}/\text{m}^3$), Aktobe ($10.6 \mu\text{g}/\text{m}^3$), Pavlodar ($13.5 \mu\text{g}/\text{m}^3$), Taldykorgan ($32.1 \mu\text{g}/\text{m}^3$), Uralsk ($36.5 \mu\text{g}/\text{m}^3$) and Ust-Kamenogorsk ($28.8 \mu\text{g}/\text{m}^3$). At the same time, Aktau, Atyrau, Taraz, Pavlodar, Ust-Kamenogorsk, Taldykorgan, Uralsk, Kyzylorda and Aktobe demonstrated a declining trend in the concentration level of $PM_{2.5}$ and PM_{10} (Tables S2, and S3 and Fig. 1). Economic growth and population growth were assumed to be the two main reasons for the increase in the pollutant concentration of the cities. Their decrease or increase would directly influence the concentrations in the future.

Some limitations were inherent to methodology. Firstly, the average annual concentrations of PM_{10} demonstrated fluctuation during 2011-2017. The quality of the monitoring system and a small number of samples could potentially be a reason. Many assumptions, particularly for the $PM_{2.5}/PM_{10}$ ratio, could also contribute to uncertainty. Moreover, other factors may contribute to PM_{10} and $PM_{2.5}$ pollution other than economic situation and demographics. According to European Environmental Agency, there are a wide range of air pollutants emitters including both man-made and natural sources, e.g., fossil fuel consumption in transport, household, electricity generation and industry. Emission from chemical and mining industries using various types of solvents in their industrial processes could be the sources. Agriculture and waste treatment by burning are also contributing to air pollution (López-Aparicio *et al.* 2013, Lucarelli *et al.* 2019). Some natural air pollution emitters are volcanoes, dust from winds, fine particles of sea-salt and plants that emit volatile organic compounds (EEA 2019). Also, urban traffic could be one of the most-influencing sources of air pollution. Coal consumption and population density rather than total population could contribute to air pollution as well (Sun *et al.* 2019, Vardoulakis and Kassomenos 2008). Meteorological conditions (precipitation, wind speed, temperature) were considered to be significant factors as well (Saramak 2019). Finally, more advanced statistical models have been currently used. Multivariate analysis of data is one of the most convenient and efficient techniques for the big data analysis. For the analysis in environmental science and management, a spatial interpolation method has been commonly used for the purpose, e.g., a geostatistical method such as ordinary kriging (Núñez-Alonso *et al.* 2019). The study has been significantly focusing on the

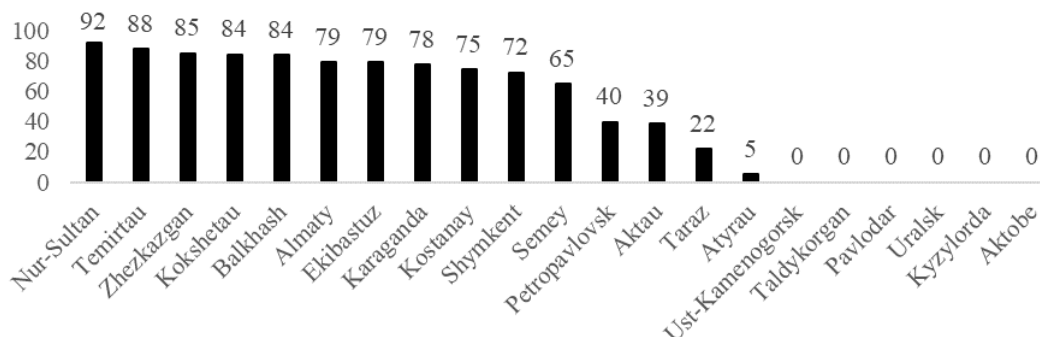


Fig. 3 Reduction in the incidence of chronic bronchitis in 2050 by the “OECD Standards” scenario in comparison to the Business as usual scenario in %

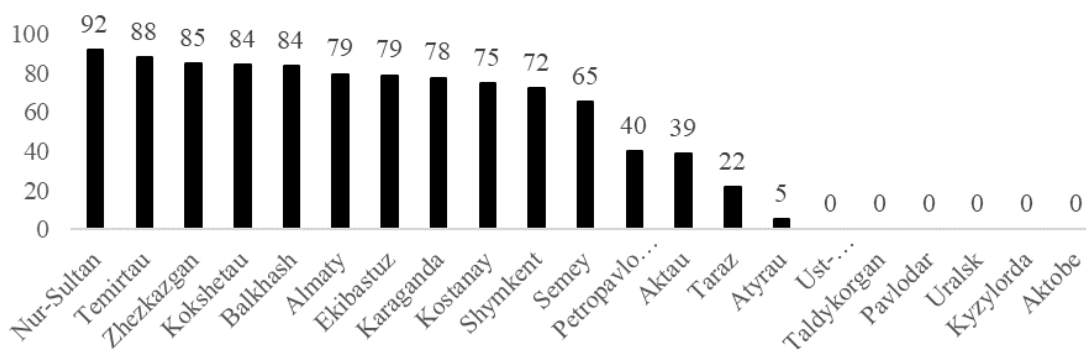


Fig. 4 Reduced days of limited activity in 2050 in the “OECD Standards” scenario in comparison to the “Business as usual” scenario in %

estimation of the monetary effect of air pollution in the country.

3.2 Impact of transition into OECD standards on mortality and diseases caused by emitted air pollutants

Fig. 2 shows the implementation of the “OECD scenario” emission restrictions leading to a reduction of mortality in a half of the cities of Kazakhstan by 2050. Reduction of the mortality in the cities of Kazakhstan reached up to 87% including Nur-Sultan (87%), Temirtau (84%), Zhezkazgan (81%), Kokshetau (81%) and Balkhash (80%). Some cities showed a minimum or zero reduction in mortality since their air quality was under or close to OECD standards (i.e., Pavlodar, Atyrau, Ust-Kamenogorsk, Uralsk, Taldykorgan, Kyzylorda and Aktobe). In general, the transition to “OECD standards” leads to 71.7% less air pollution-related mortality in 2050, i.e., 3216 less premature deaths in 2030 and 6802 in 2050 in all cities. Moreover, it leads to a 56% decrease in morbidity cases on average in 2050 (8084, 14893, 291647, 51646357, 579426, 164370146 fewer mortality cases for chronic bronchitis, number of hospitalizations, requests for urgent medical help, days of limited activity, lower respiratory ways’ disease cases in children and symptoms of respiratory disease, respectively).

Morbidity was calculated using 6 factors, but chronic bronchitis and days of limited activity were considered as the most important factors because their contribution to overall morbidity costs

Table 8 Estimated costs of morbidity and mortality associated with air pollution, \$ million

City	Baseline	Business as usual		OECD	
	2017	2030	2050	2030	2050
Nur-Sultan	146.5	304.7	1120.3	54.7	134.9
Zhezkazgan	10.7	14.6	24.4	3.0	4.3
Temirtau	22.6	33.8	62.6	6.5	9.3
Shymkent	47.0	77.2	175.1	22.2	49.1
Balkhash	6.0	9.8	20.4	2.7	3.8
Almaty	167.3	332.9	1038.6	106.3	238.4
Aktau	14.8	14.9	14.2	7.2	8.7
Atyrau	37.0	37.6	30.8	22.6	29.2
Taraz	10.2	10.1	9.1	6.1	7.1
Pavlodar	17.9	13.0	0.0	11.2	0.0
Ekibastuz	7.1	12.7	26.4	4.8	6.1
Karaganda	21.5	44.4	101.4	17.3	24.8
Semey	11.3	19.5	41.9	9.9	15.1
Ust-Kamenogorsk	9.3	10.1	8.4	10.1	8.4
Taldykorgan	3.1	3.7	4.3	3.7	4.3
Kokshetau	8.4	15.1	28.8	4.1	5.1
Petropavlovsk	2.3	5.2	11.1	5.2	6.7
Uralsk	9.2	11.9	17.2	11.9	17.2
Kyzylorda	0.0	0.0	0.0	0.0	0.0
Aktobe	0.0	0.0	0.0	0.0	0.0
Kostanay	6.8	13.2	27.4	5.9	7.3
Total	559.1	984.3	2762.5	315.3	579.8

were among the highest. Under the “OECD scenario”, the incidence of chronic bronchitis decreased to 92% in Nur-Sultan (92%), Temirtau (88%), Zhezkazgan (85%), Kokshetau (84%), and Balkhash (84%) by 2050 (Fig. 3). Overall, 3420 fewer cases of incidence chronic bronchitis in 2030, and 8084 fewer cases in 2050 occurred under the “OECD scenario”. Fig. 4 shows the changes in days of limited activity after changing emission restrictions into the “OECD Standards” scenario. Days of limited activity also decreased under the “OECD scenario” to 92% in Nur-Sultan with the same percentage difference in each city as in the case of chronic bronchitis. Similar to a reduction in mortality, four cities (i.e., Pavlodar, Ust-Kamenogorsk, Taldykorgan and Uralsk) did not experience any changes in the incidence of chronic bronchitis and the number of days of limited activity since the emission of those cities were similar or below “OECD Standards”.

To conclude, the transition into “OECD Standards” leads to positive consequences such as a substantial reduction in mortality of most of the cities, reduction in incidences of chronic bronchitis, and decreased the days of limited activity in the majority of cities. Only four cities remained at the same level regardless of the change in emission restriction standards, comprising only 14% of the total urban population of Kazakhstan. On the other hand, 86% of all cities’ population was experiencing improvement in terms of the health effect of air pollution in case of

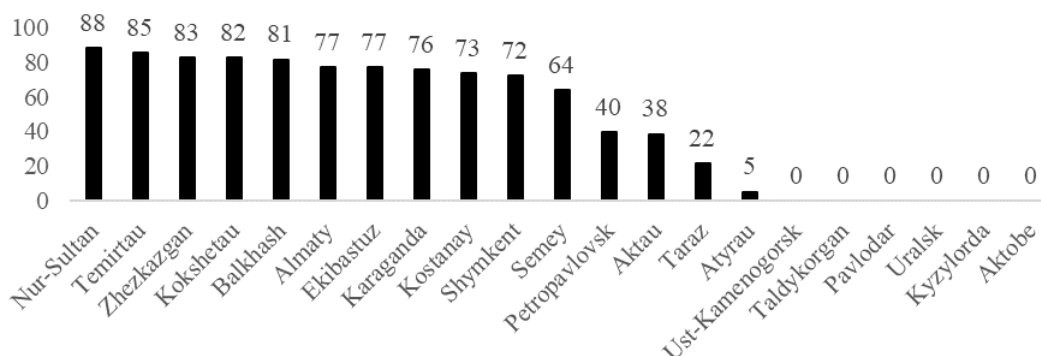


Fig. 6 Cost reduction from morbidity and mortality in 2050 in the “OECD Standards” scenario in comparison to the “Business as usual” scenario in %

transition to the “OECD” scenario. Such a significant impact on public health plays a crucial role in the economy of the country since the health care of the patients requires financial support, which is discussed in the next section.

3.3 Impact of transition into OECD standards on health care costs caused by emitted air pollutants

Table 8 illustrates the estimated costs (USD) related to incidence and premature deaths caused by exposure to air pollution in all cities of Kazakhstan in 2017, 2030 and 2050. Total cost in all cities increased from \$559 million in 2017 by 1.8 times in 2030 (\$984.3 million) and increased by 4.9 times in 2050 (\$2762.5 million) in case if the current situation persists. As can be seen from Table 8, the highest cost of morbidity and mortality is in the cities of Almaty, Nur-Sultan, Karaganda and Shymkent. This is because Almaty, Nur-Sultan and Shymkent are the cities of republican significance with a fast-growing population by > a million, while the other two cities have developed industries, i.e., Karaganda is one of the main coal producers in the country. In the case of the “OECD” scenario, the total cost in all cities increased only by 4% in 2050 (\$579.8 million) compared to 2017, while in 2030, it decreased by 43% (\$315.3 million). The cost of health care due to air pollution could significantly influence the country’s economy under assumption that the future economic growth remains at the same level and there is no change in air pollution standards. Hence, it is critical to implement the preliminary measures against current emission restriction standards to protect a significant part of the population from the possibility of mortality and morbidity by the air pollution.

In case, if emission restriction policies of 21 Kazakhstan cities could be changed to the OECD air quality standards, the economic cost of morbidity and mortality associated with air quality would decrease by \$669 million in 2030 and \$2183 million in 2050. The greatest savings in 2050 could be found to be in the cities of Nur-Sultan and Almaty, \$985 million, and \$800 million, respectively, since these cities had the largest susceptible population and high level of air pollution. On the other hand, a cost reduction relative to the 2050’s cost (as a percentage) is also important in addition to their absolute difference (Fig. 6). The highest cost reduction relative to the 2050’s cost in the “Business as usual” scenario occurs in Nur-Sultan and Temirtau, 88% and 85% respectively. Nur-Sultan is the capital of the country with one of the largest populations and air

Table 9 Comparison of the study with published works

Location	Contaminant	Air pollution reduction scenario	Reduction in mortality	Reduction in healthcare costs (USD)	Reference
Kazakhstan, 22 major cities	PM10 and PM2.5	Implementing stricter air quality standards by 2050 (OECD)	6802 in 2050 (72%)	\$2.2 billion (2050)	This study
South Africa	PM2.5	Implementing stricter air quality standards (WHO)	28000	\$29.1 billion (4.5% of South Africa's 2012 GDP)	(Altieri and Keen 2019)
Shiraz, Iran	NO ₂ , SO ₂ , PM2.5, PM10, O ₃	Addition mortalities due to difference between real concentrations and background level of 10 µg/m ³	911 cases (2016), 346 cases (2017)	Not calculated	(Bonyadi <i>et al.</i> 2020)
Beijing, China	SO ₂	Addition mortalities due to difference between real concentrations and background level of 20 µg/m ³	884 and 27854 outpatient cases	477 million RMB Yuan (2016) (~ \$3.2 billion)	(Wu <i>et al.</i> 2020)
EU	PM2.5	Achieving 50% reduction in agricultural PM2.5 emissions	~140000/year	\$407 billion /year	(Giannadaki <i>et al.</i> 2018)
Changsha, China	PM2.5	Reduction in emissions due to change in urban industrial land allocation	60.8%	\$0.69 billion	(Xu <i>et al.</i> 2020)
New York City, USA	PM2.5	Reduction in emissions due-to COVID-19 lockdowns	3455-7791	\$30.9-\$69.7 billion	(Perera <i>et al.</i> 2021)

pollution, while Temirtau is an industrial city with developed metallurgy, which causes severe air pollution problems (Abdurasulov 2018, Long 2020).

Table 9 demonstrates the comparison of the study with published works. Overall, reducing the exposure to PM10 and PM2.5 by any means leads to both public health and monetary benefits worldwide. Compared to state-scale studies (South Africa and EU) and even single city-focused study (New York City), Kazakhstan's reduction in mortality is relatively small probably due to Kazakhstan's small population (~18 million people (World Bank 2021)) or differences in the methodology. Monetary effect was approximately lower in developing countries (< \$30 billion) (Kazakhstan, China, South Africa) compared to cost reduction in developed states (USA and EU).

Additional expenditures on health care of patients due to air pollution and days of limited activity of the population would cause a significant economic impact. Considering the GDP of Kazakhstan, the cost of morbidity and mortality was 0.34% of GDP (\$166800 million (World Bank 2017)) in 2017. At the most optimistic scenario of Kazakhstan's GDP's growth (by 4.5 in 2050 (Syzdykbaev 2016)), the cost of air pollution-related health problems could constitute 0.37% of estimated GDP and increase more if the future GDP could be lower than this estimate. Thus, the cost of air pollution would increase with years, being a minimum of 0.37% of GDP and a maximum of 1.66% (assuming that the GDP would be the same in 2050 in the worst scenario). In contrast, in case of changes in air quality standards to EU air quality standards ("OECD scenario"), the economic costs of morbidity and mortality would comprise 0.35% in 2050 in the case of the

worst GDP growth scenario (same 2017's GDP in 2050) and decrease to 0.08% in the optimistic case. Thus, a change of air quality standards can lead to huge savings, which could positively influence the country's economy and decreased the level of morbidity and mortality of its citizens.

Numerous factors might affect the validity of results. Although the current investigation just focused on PM₁₀ and PM_{2.5}, there are other pollutants, which may affect mortality such as CO, NO_x, SO₂, ozone, and heavy metals (Pervin *et al.* 2008). Thus, the derived number of mortalities and costs could be further increased by adding more pollutants in the assessment. The consideration of different scenarios could further increase costs. For example, short-term acute effects in adults could be significant for the cost estimations. Moreover, other significant and costly health effects of air pollution could be overlooked. Studies are suggesting a connection of dementia and low birth weight to air pollution, for instance (Pimpin *et al.* 2018). Intangible effects such as pain and suffering have been also monetized in the literature (Pervin *et al.* 2008). The PMs in this study was derived from outdoor measurements, while people spent most of the time indoors. Indoor pollution levels are considered to be more serious and influential than that of outdoor pollution (Kumar *et al.* 2016). It is worth pointing out that indoor activities such as cooking and heating significantly deteriorate the quality of air (Pervin *et al.* 2008). In rural areas of Kazakhstan, where air pollution-producing heating with stoves is abundant, assessment should be done as well to cover all populations in the country. Thus, costs could be underestimated in the study because only outdoor effects were taken into account. Moreover, mostly long-term effects were considered, while there are also short-term costs of air pollution for adults, which may add up to the derived results. On the other hand, there are factors, which may overestimate the conclusions of the study such as using the value of a statistical life approach to estimate costs. It is designed for accidental deaths and working-age people, while air pollution affects the elderly and children at most, thereby it could lead to overestimation. Therefore, using the value of life year approach could enhance the quality of results (Delucchi *et al.* 2002, Pervin *et al.* 2008). The threshold (the PM₁₀ and PM_{2.5} concentrations level at which no health effect is observed) has no scientific derivation and different studies use different thresholds varying from 10 to 25 µg/m³ (Gao *et al.* 2015). Using a lower threshold may overestimate results because it results in a higher and overestimated difference between the current concentration and threshold.

Uncertainty is inherent to not only factors and variables but also to methodology itself. The methodology could be improved by employing stochastic methods over deterministic to show the overall picture of the distribution of costs and shed a light on the effect of parameters on the overall cost. Increasing the number of observations would also enhance the quality of results. Finally, the mortality's concentration-response factors based on epidemiological studies could contribute to uncertainty a lot. They are based on foreign populations and may not be applicable in Kazakhstan. Research on the derivation of local dose-response coefficients could help further elucidate the costs of air pollution. Moreover, it is important to note that the dose-response model was derived from a heterogeneous population. It is recommended to categorize the estimations by age, gender and income to identify vulnerable groups within the population. Further improvements for quantifying the air pollution effect may include the risk assessment procedures for urban populations. Brody and Golub (2014) suggested to establish a risk assessment institute and to install proper modern monitoring and information systems, which can produce high-quality air pollution data to obtain a better quality estimations. Overall, the study demonstrated the negative effect of the current air quality regulations in Kazakhstan and highlights the need to alter environmental policy and switch to European guidelines. This study is significant for policymaking and the enhancement of environmental regulations in Kazakhstan.

4. Conclusions

Most of the Kazakhstan cities do not meet the WHO standards of PM₁₀ concentration. The only 6 cities (Kyzylorda, Aktobe, Pavlodar, Taldykorgan, Uralsk and Ust-Kamenogorsk) would meet the environmental condition of PM₁₀ lower than the WHO limit by 2050. The incidences and deaths caused by ambient air pollution would potentially lead to an increase of economic costs by 2050 (\$559.1 million in 2017 to \$2762.5 billion by 2050). Among all cities, the highest mortality reduction was found in Nur-Sultan, Temirtau and Zhezkazgan. Nur-Sultan is the Kazakhstan capital of Republican significance with the highest rate of development, while Temirtau and Zhezkazgan are the cities with developed coal and copper mining industries. All predicted negative consequences of ambient air pollution make it essential to change the current environmental policy of the country, which could effectively decrease the total losses. Changes in the limit of PM₁₀ concentration in the environmental policy of Kazakhstan to that in the OECD standards could directly lead to a decrease in the losses of the country's economy and citizens' health. The cost related to morbidity and mortality by the air pollution would possibly decrease by \$669 million in 2030 and \$2,183 million in 2050 with the greatest saving in Almaty and Nur-Sultan.

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Appendix



Fig. S1 Cities of Kazakhstan analyzed in the study

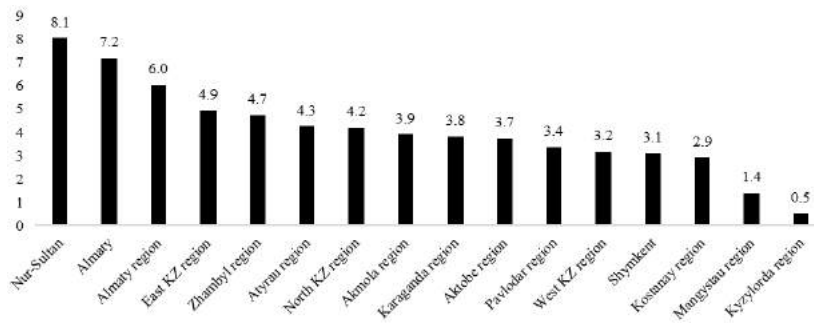


Fig. S2 Region rating by average annual GRP growth for the period of 2011-2017 (Committee on Statistics 2018)

In 2017, Shymkent was a part of South Kazakhstan region but now it is separate city. GRP value demonstrated here accounts for modern Shymkent city and Turkestan region together.

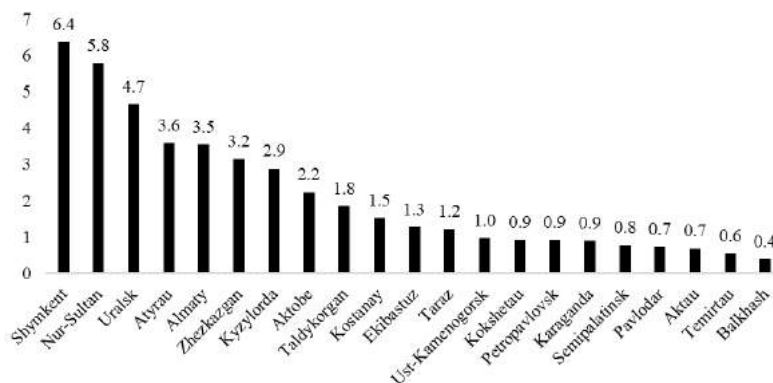


Fig. S3 City rating by average annual population growth for the period of 2011-2017 (Committee on Statistics 2018)

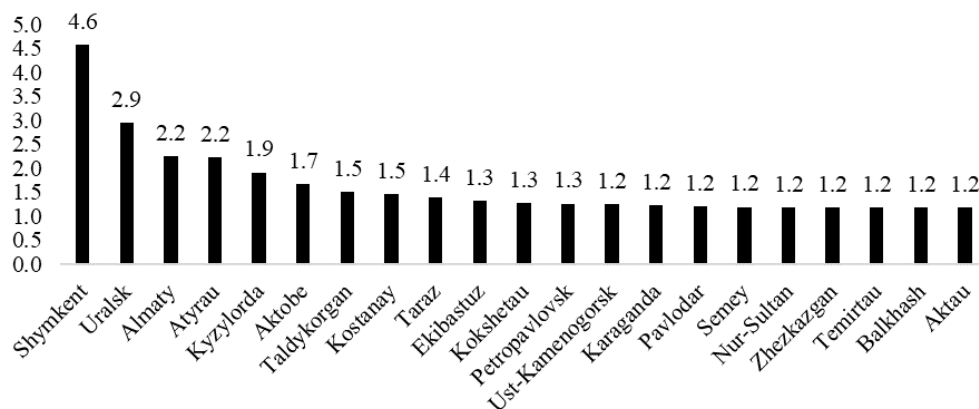


Fig. S4 Forecast values for the population in 2050 compared to 2017

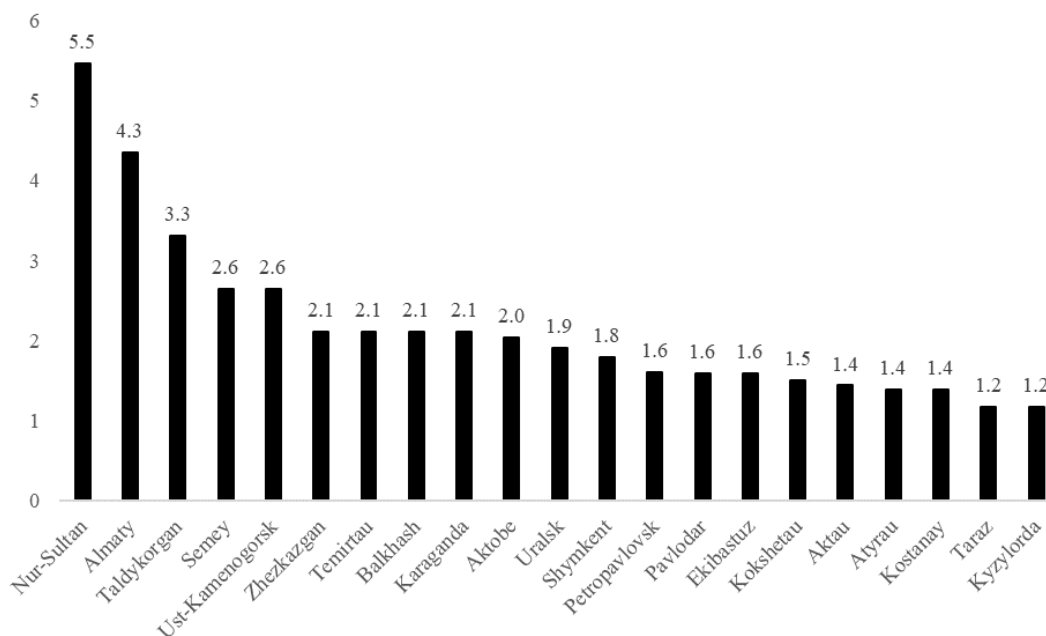


Fig. S5 Forecast values for the gross regional product (GRP) in 2050 compared to 2017

Table S1 Location of observation posts

City	Frequency	Type of sampling	Location
Nur-Sultan	3 times/day	Manual	Dzhambul st. 211
	3 times/day	Manual	Auezov-Seyfullin st. intersection
	3 times/day	Manual	Tashkent st., Lesozavod area
	3 times/day	Manual	Valikhanov st.-Bogembay Batyr ave. intersection, "Shapagat" market
Zhezkazgan	3 times/day	Manual	Saryarka st., knitwear fabric area
	3 times/day	Manual	Zhastar st. 6, Metallurg square

Table S1 Continued

	3 times/day	Manual	Dmitrov st. 212
Temirtau	3 times/day	Manual	6th district “Amangeldy”/ Temirtau st.
	3 times/day	Manual	3a district (near emergency station)
	3 times/day	Manual	Abay ave., JSC “Yuzhpolymetal”
Shymkent	3 times/day	Manual	Ordabassy square, Kazybek bi st. and Tole bi st. intersection
	3 times/day	Manual	Aldiyarov st. 6, JSC “Shymkentcement”
	3 times/day	Manual	Sairam st. 198, near brewery
	3 times/day	Manual	Sabitov district (near School #6)
Balkhash	3 times/day	Manual	Lenin st. and Alimzhanov st. intersection
	3 times/day	Manual	Kirov st. (near hospital)
	3 times/day	Manual	Amangeldy st. and Satpayev st. intersection
Almaty	4 times/day	Manual	Amangeldy st. and Satpayev st. intersection
	3 times/day	Manual	Rayimbek ave. and Nauryzbay batyr st. intersection
	3 times/day	Manual	Ainabulak district 3
	3 times/day	Manual	Marechek st. and Momyshuly st. intersection
	3 times/day	Manual	Tastak district 1, Tole bi st. 249
Aktau	3 times/day	Manual	1st district, Caspian department of ecology
	3 times/day	Manual	Aktau seaport
Atyrau	3 times/day	Manual	Azattyq ave. and Auezov ave. intersection
	3 times/day	Manual	Satpayev ave. and Vladimirskaya st. intersection
Taraz	3 times/day	Manual	Shymkent st. 22
	3 times/day	Manual	Rysbek batyr st. 15 and Niyetkaliyev st. intersection
	3 times/day	Manual	Abay st. and Tolebi st. intersection
	3 times/day	Manual	Bayzak batyr st. 162
Pavlodar	3 times/day	Manual	Kamzinn st. and Chkalov st. intersection
	3 times/day	Manual	Aimanov st. 26
Ekibastuz	3 times/day	Manual	Berkembayev st and Satpayev st. intersection, 8th district
Karaganda	4 times/day	Manual	Airport area
	3 times/day	Manual	Lenin st. and Bukhar Zhyrau ave. 1 intersection
	3 times/day	Manual	Birysov st. 15
	3 times/day	Manual	Ermekov st. 116
Semey	3 times/day	Manual	Ryskulov st. and Glinka st. intersection
	3 times/day	Manual	343th district (kindergarden)
Ust-Kamenogorsk	3 times/day	Manual	Rabochaya st. 6
	3 times/day	Manual	Kaysenov st. 30
	3 times/day	Manual	1st October st. 126
	3 times/day	Manual	Egorov st. 6
	3 times/day	Manual	Satpayev ave. 12
Taldykorgan	3 times/day	Manual	Gagarin st. 216 and Zhabayev st. intersection
Kokshetau	3 times/day	Manual	Meteostation

Table S1 Continued

Petropavlovsk	3 times/day	Manual	Valikhanov st. 17
	3 times/day	Manual	Buketov st. 16 and Kazpravda st. intersection
Uralsk	every 20 min	Automatic	Gagarin st. 25
	every 20 min	Automatic	Daumov st. (near Kirov park)
	every 20 min	Automatic	Muhit st. (Mirlan market)
Kyzylorda	3 times/day	Manual	Muratbayev st. 24-a
Aktobe	4 times/day	Manual	Aviagorodok 14
	3 times/day	Manual	Belinskii st. 5
	3 times/day	Manual	Lmonosov st. 7
Kostanay	3 times/day	Manual	Kairbekov st. 379
	3 times/day	Manual	Doschanov st. 43
	every 20 min	Automatic	Borodin st.
	every 20 min	Automatic	Mayakovskii st.

Table S2 Predicted concentrations of PM₁₀ in the OECD Standards scenario, µg/m³

City	2017 ¹	2020	2025	2030	2035	2040	2045	2050
Astana	138	105	73	40	40	40	40	40
Zhezkazgan	135	104	72	40	40	40	40	40
Temirtau	131	101	70	40	40	40	40	40
Shymkent	103	82	61	40	40	40	40	40
Balkhash	87	71	56	40	40	40	40	40
Almaty	82	68	54	40	40	40	40	40
Aktau	74	63	51	40	40	40	40	40
Atyrau	64	56	48	40	40	40	40	40
Taraz	62	54	47	40	40	40	40	40
Pavlodar	61	54	47	40	37	29	22	13
Ekibastuz	59	53	46	40	40	40	40	40
Karaganda	54	49	45	40	40	40	40	40
Semipalatinsk	52	48	44	40	40	40	40	40
Ust-Kamenogorsk	45	44	42	40	38	35	32	29
Taldykorgan	42	41	40	39	37	36	34	32
Kokshetau	74	63	51	40	40	40	40	40
Petropavlovsk	28	32	35	39	40	40	40	40
Uralsk	41	41	40	40	39	38	37	36
Kyzylorda	15	14	13	12	11	9	8	6
Aktobe	12	12	12	12	11	11	11	11
Kostanai ²	48	46	43	40	40	40	40	40

¹Average concentration during 2015-2017²Measured value of PM₁₀

Table S3 Predicted concentrations of PM_{2.5} in the air up to 2050 according to the “Business as usual” scenario, µg/m³

City	2017	2020	2025	2030	2035	2040	2045	2050
Astana	62.1	65.8	72.1	80.3	91.0	104.7	122.6	145.6
Zhezkazgan	60.9	62.4	64.8	67.5	70.5	73.9	77.7	82.0
Temirtau	58.9	61.8	66.8	72.2	78.1	84.6	91.7	99.5
Shymkent	46.5	46.5	46.6	46.8	46.9	47.1	47.2	47.4
Balkhash	39.0	41.5	45.8	50.6	55.9	61.9	68.6	76.2
Almaty	37.1	38.5	40.9	43.7	47.1	51.1	55.8	61.3
Aktau	33.4	32.7	31.5	30.3	29.0	27.7	26.4	25.1
Atyrau	28.7	27.9	26.9	25.6	24.2	22.6	20.7	18.6
Taraz	27.8	27.4	26.4	25.5	24.4	23.4	22.3	21.1
Pavlodar	27.6	25.9	22.9	19.8	16.6	13.2	9.7	6.1
Ekibastuz	26.5	28.8	33.1	37.7	42.7	48.0	53.8	60.0
Karaganda	24.1	27.7	31.9	36.3	41.0	46.0	51.2	56.8
Semipalatinsk	23.3	24.6	26.9	29.2	31.5	34.0	36.4	38.9
Ust-Kamenogorsk	20.2	19.7	18.9	18.0	17.0	15.8	14.5	13.0
Taldykorgan	19.0	18.6	18.1	17.5	16.8	16.1	15.3	14.5
Kokshetau	33.2	37.4	43.5	49.8	56.4	63.3	70.5	78.0
Petropavlovsk	12.4	14.2	15.9	17.7	19.5	21.4	23.4	25.5
Uralsk	18.4	18.3	18.0	17.8	17.5	17.2	16.8	16.4
Kyzylorda	6.7	6.4	5.9	5.4	4.8	4.2	3.5	2.7
Aktobe	5.5	5.5	5.4	5.3	5.2	5.0	4.9	4.8
Kostanay	62.1	65.8	72.1	80.3	91.0	104.7	122.6	145.6